

Screening of some filamentous fungi for cadmium tolerance in aquatic environments

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Abstract: Heavy metals (HMs), such as cadmium (Cd), ingress into the human body via the food chain through animal and plant agrochemicals, inducing epigenetic modifications, DNA damage, genetic mutations, and carcinogenesis. This investigation aimed to scrutinize and identify fungi tolerant to Cd toxicity in aqueous solutions using the poisoned food technique. Thirty fungal strains were cultured under three distinct concentrations of Cd (0, 100, and 300 mg/L), with subsequent quantitative evaluation of mycelial growth. Cluster analysis delineated two fungal groups comprising 19 and 11 strains. Group II demonstrated superior performance across most evaluated traits compared to Group I, with exceptions noted for biomass at 0 mg/L and stress tolerance index (STI) at 300 mg/L Cd. Consequently, Group II was deemed the superior cohort. Within Group II, Epicocum nigrum exhibited the highest biomass production (2.17 and 1.35 g/L, respectively) at both 0 and 100 mg/L Cd concentrations, alongside the highest STI (2.18 and 0.43, respectively) at Cd levels of 100 and 300 mg/L. Conversely, Clonostachys displayed the highest rogersoniana biomass production (0.46 g/L) at a Cd concentration of 300 mg/L, coupled with the lowest percentage of inhibition (PI) (-11.35 and -31.40%, respectively) at Cd toxicity levels of 100 and 300 mg/L. Hence, the 11 strains within Group II, particularly E. nigrum and C. rogersoniana, exhibit promise for further

Submitted 15 May 2024, accepted for publication 19 June 2024 Corresponding Author: E-mail: m.tajick@gmail.com © 2024, Published by the Iranian Mycological Society https://mij.areeo.ac.ir investigation concerning their efficacy in Cd removal from aqueous solutions.

Keywords: Cluster analysis, fungi, heavy metal toxicity, inhibition rate, stress tolerance index.

INTRODUCTION

Environmental pollution presents a significant global health challenge. The challenges posed by HM pollutants are further compounded by diverse industrial and agricultural activities on a worldwide scale (Xu et al., 2011). Effective strategies are imperative for removing HMs, either through separation and concentration or recovery and reuse. Among HMs, Cd, Ni, Cr, Pb, Hg, As, Fe, Cu, and Zn are harmful to humans due to their high toxicity, tendency to accumulate, and stability (Sud et al., 2008).

Cadmium (Cd), renowned for its high toxicity and mobility among HMs, is a by-product of Zn mining, smelting, and refining. Its exceptional significance arises from being 20 times more toxic to plants compared to the other metals. It exhibits an estimated average biological half-life of 18 years in the environment and 10 years within the human body (Salt et al., 1995). Extensive agricultural soils worldwide are contaminated with low to moderate Cd concentrations due to prolonged usage of phosphate fertilizers and sewage sludge (Vassilev et al., 2003). The typical Cd concentration in soil is approximately 1 μ g/g, with toxicity thresholds ranging from 3 to 8 μ g/g and critical limits in plants ranging from 5 to 30 μ g/g (Kubier et al., 2019). HMs such as Cd permeate the human body through the food chain via animal and plant agrochemicals, leading to epigenetic alterations, DNA damage, genetic mutations, and carcinogenesis (Knasmüller et al., 1998). The human threshold for Cd absorption is 0.1 mg/kg per day (Satarug et al., 2017), with associated risks including kidney failure, hypertension, hepatic and pulmonary damage, bronchitis, atherosclerosis, genetic mutations, and cancer (Brahman et al., 2013; Davis et al., 2003).

Various methods including chemical, physical, electrochemical, and biological approaches have been employed to remove HM (Imre-Lucaci et al., 2011). Presently, biological absorption stands out as the most practical and effective method for decontaminating HM-laden wastewater and reducing metal Biological concentrations therein. absorption encompasses the uptake of ions by microorganisms, encompassing microbial leaching, bioaccumulation, and enzymatic bioabsorption. transformation (Soleimanifar et al., 2012). Microbial mechanisms involved in HM bioremediation play pivotal roles in anaerobic respiration reduction, detoxification, uptake, dissolution, and accumulation (Siddiquee et al., 2015). Genetic engineering and chemical modification enhance HM removal by modifying cell surface components and augmenting microorganism adsorption capacities (Zafar et al., 2007).

Microorganisms with high HM tolerance are typically selected for bioremediation, with numerous microbial species, including algae, bacteria, yeasts, and fungi, serving as bio-based adsorbents with significant absorption potential (Mungasavalli et al., 2007). Fungi, owing to their distinctive properties such as cation and anion binding to fungal cell walls, absorption exhibit superior HM capabilities (Chojnacka, 2010). Consequently, both living and deceased fungal cells possess notable capacities for absorbing toxic metals from aqueous solutions (Wang & Chen, 2009) and are frequently employed in HM recovery processes (Kacprzak & Malina, 2005). Notably, Penicillium, Aspergillus, and Rhizopus fungi have been extensively researched for HM removal and biotreatment in contaminated environments (Gomes et al., 2014; Rezaei et al., 2022; Bahari Saravi et al., 2022). For instance, a study explored the efficacy of three Trichoderma species (T. asperellum, T. harzianum, and T. tomentosum) in removing Cd ions, indicating significant potential for Cd ion reduction and bioremediation across varied pH and concentration conditions (Mohsenzadeh & Shahrokhi, 2014). Additionally, Yaghoubian et al. (2019) investigated the biological removal of Cd from aqueous solutions using filamentous fungi (Piriformospora indica and six strains of Trichoderma species), highlighting the high Cd tolerance and absorption capacity of Trichoderma spp., particularly T. simmonsii. The researchers suggested Trichoderma species, especially Τ. simmonsii (UTFC 10063), as potent bio-remediation agents in Cd-contaminated aqueous solutions (Yaghoubian et al., 2019). Furthermore, the bioremediation potential of Phlebia brevispora in Cd, and Ni was examined, removing Pb, demonstrating applicability for its HM bioremediation in industrial wastewater (Sharma et al., 2020). Previous studies have underscored the influence of various factors, including species type, HM type and concentration, pH, temperature, and biosorbent nature, fungi-mediated HM on bioremediation (Zaghir & Kakhki, 2021). For example, recent research revealed significant increases in residual metal concentrations and absorption percentages at all levels by altering Cd and Ni concentrations (Zaghir & Kakhki, 2021). The exploration of novel fungal species holds immense potential for enhancing the efficacy, specificity, and sustainability of bioremediation strategies for HM removal from contaminated environments. This study aims to screen and identify fungi tolerant to cadmium toxicity in aqueous solutions.

MATERIALS AND METHODS

To investigate the tolerance of 30 fungal strains to Cd toxicity, a laboratory experiment was conducted at Sari Agricultural Sciences and Natural Resources University (SANRU), Mazandaran, Iran.

Fungal strains

The fungal strains used in this study were obtained from the Department of Plant Pathology and Genetics and the Agricultural Biotechnology Institute of Tabarestan (GABIT), SANRU, Sari, Iran. The strain names used are listed in Table 1. Fungal strains were grown on the sterile potato dextrose agar (PDA, Merck) medium at 28°C for 10 d.

Fungal strains screening

To screen fungal strains for tolerance to Cd toxicity and to determine the appropriate concentration of Cd, 30 strains of different fungi were cultured in three concentrations of Cd (0, 100, and 300 mg/L) from the source of Cd chloride (CdCl₂•2.5H₂O). The Cd tolerance of the fungi was quantitatively assessed using the poisoned food technique (Dhingra and Sinclair, 1995). Potato dextrose broth (PDB, Merck) medium was prepared and distributed in a volume of 25 ml in 50 ml sterilized Erlenmever flasks. The desired concentration of Cd was achieved by adding appropriate quantities of CdCl₂•2.5H₂O stock solution to the culture medium. The flasks were inoculated with 4-mm agar discs of 10-day-old cultures of fungal strains. The samples were then placed in a shaking incubator for 15 days at 28°C and a rotation speed of 120 rpm (Yaghoubian et al., 2019). Three replicate flasks were used per concentration. Then, the dry weight, inhibition rate, and stress tolerance index (STI) of the fungal strains at different Cd concentrations were assessed.

Determination of biomass

Mycelial biomass (g/L of the culture medium) was used as a measure of fungal growth. After 15 days of incubation, the fungi were examined by measuring the dry weight of the biomass. To achieve this, biomass was harvested by filtering through filter paper (Whatman no.42), washed three times with deionized water, and incubated at 65 °C for 24 h. The mycelial dry weight was measured using a digital scale with a sensitivity of 0.0001.

Calculation of fungal growth inhibition rate

The growth inhibition rate was expressed as the percentage reduction in fungal biomass compared with that in the control group. The inhibition rate was calculated using the equation described by Yaghoubian et al. (2019) (Equation 1), where the percentage of inhibition (PI) was determined using the fungal biomass (mg/L) in the control (0 mg/L Cd), culture media (C) and the fungal biomass (mg/L) in the Cd-amended culture media (T):

$$PI = \left[\frac{(C-T)}{C}\right] \times 100 \tag{1}$$

Calculation of stress tolerance index

The stress tolerance index (STI) was calculated based on the dry biomass of the fungal strains under both stress and control conditions, utilizing the following equation:

$$STI = \frac{(Yn \times Ys)}{(\bar{Y}n)^2}$$
(2)

where Yn, biomass of fungal strains in control group (mg/L); Ys; biomass of fungal strains under Cd toxicity (mg/L); and $\bar{Y}n$, average biomass of all fungal strains under control conditions (Wen et al, 2023).

Statistical analysis

The obtained data were analyzed using the oneway ANOVA and means of treatments were compared by LSD test at the $P \le 0.05$ level by statistical analysis system (SAS) version 9.4. Based on the traits, a hierarchical clustering analysis using Euclidean distances and Ward's minimum variance method was used.

In this study, various fungi were cultured across three Cd concentrations: 0, 100, and 300 mg/L. Sari Agricultural Sciences and Natural Resources University (SANRU), Sari, Iran.

RESULTS

In this study, various fungi were cultured across three Cd concentrations: 0, 100, and 300 mg/L. These microorganisms exhibited significant tolerance, surviving even at elevated Cd concentrations. However, distinct growth patterns and tolerance levels were observed among different species (Fig 1 & Table 2).

B. dotidae showed a notable decrease (94%) in biomass as Cd concentration increased to 300 mg/L.

C. rogersoniana exhibited a unique response, with a noticeable biomass increase (37%) from 0 to 300 mg/L Cd, albeit lower than most isolates under control conditions. *P. lilacinus* recorded maximum biomass (4.891 g/L) at 0 mg/L Cd but decreased at 100 mg/L (0.247 g/L) before slightly increasing at 300 mg/L (0.453 g/L). *B. dotidae* initially decreased in growth at 100 mg/L Cd but later increased at 300 mg/L.

B. dotidae exhibited the highest inhibition at Cd concentrations of 100 mg/L (98.11%) and 300 mg/L (94.64%). In contrast, C. rogersoniana displayed the (-11.35%) lowest inhibition and -31.40%. respectively). P. lilacinus and T. lixii were identified as the most resistant and sensitive to Cd exposure, respectively (Table 2). After clustering analysis at 100 and 300 mg/L Cd levels, two distinct groups emerged: Group I comprising 19 fungal strains, and Group II comprising 11 fungal strains (Fig 2). Wilks' Lambda test confirmed significant differences between these groups (P < 0.001), with a grouping accuracy of 96.7% (Table 3).

In inter-group comparisons, Group I produced more biomass under Cd-free conditions but exhibited higher inhibition rates when exposed to Cd (PI = 74.33% and 81.01% at 100 and 300 mg/L, respectively). Conversely, Group II displayed higher biomass production in the presence of Cd and a higher STI at 100 mg/L Cd. *E. nigrum* in Group II showed the highest biomass production and STI at both 0 and 100 mg/L Cd concentrations (Table 4). *C. rogersoniana* exhibited the highest biomass production at 300 mg/L Cd and the lowest PI at both Cd toxicity levels (Table 2).

This comprehensive analysis underscores the diverse responses of fungal strains to Cd toxicity, critical for understanding their potential applications in bioremediation or industrial processes.

DISCUSSION

Thirty strains of various filamentous fungi underwent evaluation for biomass production under cadmium (Cd) toxicity utilizing the poisoned food technique. Our findings indicate a decrease in biomass production across all fungi with increasing Cd concentrations, highlighting species-specific tolerance variations to Cd toxicity. Similar observations were made by Liaquat et al. (2020), who investigated the growth of isolated fungi under elevated concentrations of Cd, Cr, and Pb. Previous studies have also documented the adverse effects of Cd and other HMs on fungal biomass production. However, numerous studies have demonstrated fungi's capability to adapt and flourish in high metal concentrations.

Cluster analysis identified two distinct groups of fungi, with Group II exhibiting superior performance across most examined traits, except for biomass under Cd-free conditions and stress tolerance index (STI) at 300 mg/L Cd concentration. Consequently, Group II was deemed the superior group.

Within the 11 fungal strains comprising Group II, *Epicoccum nigrum* displayed the highest biomass production at both 0 and 100 mg/L Cd concentrations,

along with the highest STI when exposed to Cd levels of 100 and 300 mg/L. Conversely, *Clonostachys rogersoniana* exhibited the highest biomass production at a Cd concentration of 300 mg/L, concurrently displaying the lowest phytotoxicity index (PI) at both Cd toxicity levels.

Table 1. List of fungal strains examined in this study, sourced from the Department of Plant Pathology and the Agricultural Biotechnology Institute of Tabarestan (GABIT), Sari Agricultural Sciences and Natural Resources University (SANRU), Sari, Iran.

NO	Fungi strains name	NO	Fungi strains name
1	Acrostalagmus luteoalbus	16	Paecilomyces marquandii
2	Alternaria tenuissima	17	Paecilomyces lilacinus
3	Aspergillus terreus	18	Paraconithyrium brasiliense
4	Botyophaeria dotidae	19	Penicilium pinophylium
5	Ceriporia alachuana	20	Phoma medicagenis
6	Cladosporium perangustum	21	Phoma rabiei
7	Clonostachys rogersoniana	22	Phomopis nobolis
8	Epicocum nigrum	23	Scytalidium thermophilum
9	Fusarium oxysporum	24	Trichoderma citrinoviride
10	Hypoxylon investines	25	Trichoderma harzianum
11	Hypoxylon sabmonticlosum	26	Trichoderma lixii
12	Lentisnigiacnus tigrinus	27	Trichoderma logibrachiatum
13	Metarhizium anisopliae	28	Trichoderma simmonsii (UTFC10061)
14	Montagnula sp.	29	Trichoderma simmonsii (UTFC10063)
15	Nigrospora sphaeria	30	Xylaria sinerea



Fig 1. Growth pattern of fungal strains (A: *P. marquandii*, B: *P. lilacinus*, C: *C. rogersoniana*, D: *P. brasiliense*, E: *E. nigrum*, F: *H. investines*, G: *T. longibrachiatum*, H: *T. simmonsii*) in different cadmium concentration (0, 100, and 300 mg/l) in broth culture, 15 days after inoculation.

Table 2. Effect of cadmium	concentrations or	ı biomass	production,	percentage	of inhibition	(PI) and	stress	tolerance
index (STI) of fungal strains	in the broth cultu	re.						

Cluster		Dry weight (g/l)			PI ((%)	STI	
groups	Fungi strain	(admium (m	ig/l)		
		0	100	300	100	300	100	300
Ι	Scytalidium thermophilum	0.515 ^{ijk}	0.176 ^{de}	0.135 ^{e-j}	65.40 ^{a-h}	75.59 ^{a-f}	0.070 ^{cd}	0.058 ^{hij}
	Trichoderma lixii	0.291 ^k	0.104 ^e	0.070 ^j	64.16 ^{a-h}	76.06 ^{a-f}	0.022 ^{cd}	0.015 ^j
	Lentisnigiacnus tigrinus	0.841^{hij}	0.227 ^{de}	0.131 ^{e-j}	62.60 ^{a-h}	83.30 ^{a-d}	0.119 ^{cd}	0.089 ^{g-j}
	Trichoderma simmonsii (UTFC10063)	0.528 ^{ijk}	0.202 ^{de}	0.105^{hij}	60.39 ^{b-h}	79.06 ^{a-f}	0.076 ^{cd}	0.039 ^{ij}
	Xylaria sinerea	1.210^{fgh}	0.509^{bcd}	0.183 ^{d-j}	55.26 ^{c-i}	82.64 ^{a-d}	0.474^{bcd}	0.158 ^{e-i}
	Hypoxylon investines	0.873^{ghi}	0.344^{cde}	0.124 ^{f-j}	58.68 ^{c-h}	86.00 ^{a-d}	0.211 ^{cd}	0.081^{hij}
	Aspergillus terreus	2.118 ^{bcd}	0.646 ^{bc}	0.179 ^{d-j}	69.24 ^{a-g}	91.59 ^{ab}	1.001 ^b	0.281 ^{cde}
	Penicilium pinophylium	0.444^{ijk}	0.075 ^e	0.111 ^{g-j}	71.95 ^{a-g}	67.21 ^{a-f}	0.015 ^d	0.033 ^{ij}
	Trichoderma logibrachiatum	0.457^{ijk}	0.124 ^e	0.169 ^{d-j}	73.06 ^{a-f}	63.75 ^{c-g}	0.043 ^{cd}	0.059^{hij}
	Cladosporium perangustum	1.159 ^{fgh}	0.337 ^{cde}	0.259 ^{b-f}	71.06 ^{a-g}	77.50 ^{a-f}	0.290 ^{cd}	0.220 ^{d-g}
	Trichoderma simmonsii (UTFC10061)	0.387 ^{jk}	0.101°	0.093 ^{ij}	73.29 ^{a-f}	75.75 ^{a-f}	0.028 ^{cd}	0.026 ^{ij}
	Nigrospora sphaeria	0.532 ^{ijk}	0.201 ^{de}	0.145 ^{d-j}	70.05 ^{a-g}	72.61 ^{a-f}	0.104 ^{cd}	0.065^{hij}
	Ceriporia alachuana	1.306 ^{fg}	0.264^{de}	0.242 ^{b-g}	76.95 ^{a-e}	79.65 ^{a-e}	0.240 ^{cd}	0.230^{def}
	Phoma medicagenis	1.264^{fgh}	0.269 ^{de}	0.344 ^{ab}	78.41 ^{a-e}	72.95 ^{a-f}	0.247 ^{cd}	0.322^{bcd}
	Phomopis nobolis	2.520 ^b	0.313^{cde}	0.233 ^{b-h}	87.00 ^{a-d}	90.34 ^{abc}	0.558 ^{bc}	0.415 ^{bc}
	Phoma rabiei	1.828 ^{de}	0.182^{de}	0.137 ^{d-j}	90.30 ^{abc}	92.64ª	0.252 ^{cd}	0.188^{d-h}
	Alternaria tenuissima	2.235 ^{bcd}	0.184^{de}	0.264 ^{b-e}	91.20 ^{abc}	87.29 ^{a-d}	0.289 ^{cd}	0.412 ^{bc}
	Paecilomyces lilacinus	4.891 ^a	0.247 ^{de}	0.453ª	95.05 ^{ab}	90.75 ^{abc}	0.911 ^b	1.638ª
	Botyophaeria dotidae	2.388 ^{bc}	0.083°	0.137 ^{d-j}	98.11ª	94.64ª	0.067 ^{cd}	0.256 ^{de}
II	Paecilomyces marquandii	0.307 ^k	0.277 ^{de}	0.196 ^{d-j}	10.80 ^{jkl}	36.50 ^g	0.063 ^{cd}	0.045 ^{ij}
	Fusarium oxysporum	0.205 ^k	0.186 ^{de}	0.119 ^{g-j}	19.14 ⁱ⁻¹	39.30 ^g	0.034 ^{cd}	0.019 ^j
	Hypoxylon sabmonticlosum	0.366 ^k	0.356 ^{cde}	0.164 ^{d-j}	1.10 ^{kl}	64.14 ^{b-g}	0.104 ^{cd}	0.057^{hij}
	Trichoderma citrinoviride	0.449 ^{ijk}	0.201 ^{de}	0.192 ^{d-j}	46.81 ^{e-j}	51.51 ^{fg}	0.059 ^{cd}	0.060^{hij}
	Trichoderma harzianum	0.478^{ijk}	0.206 ^{de}	0.208 ^{c-i}	53.75 ^{d-i}	51.56 ^{fg}	0.074^{cd}	0.071^{hij}
	Acrostalagmus luteoalbus	1.369 ^{ef}	0.319 ^{cde}	0.198 ^{d-j}	51.95 ^{d-i}	60.35 ^{d-g}	0.285 ^{cd}	0.103 ^{f-j}
	Paraconithyrium brasiliense	1.218^{fgh}	0.429 ^{b-d}	0.334 ^{abc}	36.59 ^{g-k}	54.05 ^{efg}	0.265 ^{cd}	0.234 ^{def}
	Metarhizium anisopliae	1.954 ^{cd}	0.793 ^b	0.198 ^{d-j}	45.65 ^{e-j}	88.10 ^{abc}	0.913 ^b	0.270^{de}
	Epicocum nigrum	2.172 ^{bcd}	1.353ª	0.271 ^{bcd}	37.95 ^{f-j}	87.51 ^{a-d}	2.176 ^a	0.432 ^b
	Montagnula sp.	0.303 ^k	0.210 ^{de}	0.079^{ij}	31.94 ^{h-k}	71.50 ^{a-f}	0.048 ^{cd}	0.016 ^j
	Clonostachys rogersoniana	0.336 ^k	0.381 ^{cde}	0.459ª	-11.35 ¹	-31.40 ^h	0.097 ^{cd}	0.119 ^{f-j}
Signific	ant	**	**	**	**	**	**	**

Significant************** Significant at the 1% probability levels. The same letter in each column indicates non-significant different according to
LSD at 5% of probability.

Table 3. Detection function test for fungal strains under different concentrations of Cd by using Wilks' La	ambda.
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Test of Function(s)	Wilks' Lambda	Chi-square	df	Sig.
Ι	0.290	30.327	7	0.000



Fig. 2. The dendrogram obtained from the cluster analysis of fungal strains at 100 and 300 mg/L levels of Cd by using *Ward's minimum variance method*.

~	Dry weight (g/l)		PI (%)		STI		
Groups	0	100	300	100	300	100	300
Ι	1.357ª	0.241 ^b	0.185 ^b	74.327 a	81.013 a	0.264 ^b	0.241ª
Π	0.832 ^b	0.428ª	0.220ª	29.483 b	52.103 b	0.374 ª	0.130 b
Sig.	**	*	ns	**	*	*	*

Table 4. Inter-group mean comparison of growth and tolerance traits of fungal strains under different concentrations of Cd (mg/l).

ns, * and **, non-significant and significant at 5% and 1% probability level, respectively.

In each column, the means with the same letter (s) did not differ significantly based on the LSD test at the 5% level of probability.

These findings suggest that the 11 strains in Group II, particularly *E. nigrum* and *C. rogersoniana*, merit further investigation regarding their potential in Cd removal from aqueous solutions. However, studies by Zafar et al. (2007) and Yagoubian et al. (2019) have indicated that there may not be a direct correlation between metal tolerance and the bioabsorption capacity of fungal isolates. Consequently, additional research is needed to explore this aspect comprehensively.

ACKNOWLEDGMENTS

Deep gratitude is extended to the Department of Plant Pathology and the Genetics and Agricultural Biotechnology Institute of Tabarestan (GABIT) for generously providing the fungal strains. Additionally, sincere appreciation is expressed to Dr. M. Najafi for invaluable technical assistance.

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غربالگری قارچهای رشتهای متحمل به فلز سنگین کادمیوم در محیطهای آبی

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چکیده: فلزات سنگین مانند کادمیوم (Cd) از طریق زنجیره غذایی وارد بدن انسان میشوند و باعث تغییرات اپیژنتیک، آسیب به .DNA جهشهای ژنتیکی و سرطان زایی میشوند. این پژوهش با هدف غربالگری و شناسایی قارچهای متحمل به سمیت کادمیوم در محلولهای آبی با استفاده از روش غذای مسموم اجرا شد. بنابراین، ۳۰ جدایه قارچ در سه غلظت متفاوت کادمیوم (۰، ۱۰۰، ۲۰۰۰ میلی گرم در لیتر) کشت شده و میزان رشد میسیلیوم به صورت کمی مورد ارزیابی قرار گرفت. در تجزیـه و تحلیـل خوشـهای (کلاستر)، دو گروه قارچ بهترتیب با ۱۹ و ۱۱ جدایه شناسایی شدند. گروه دوم در بیشتر صفات مورد بررسی، بهجز زیستتوده در غلظت صفر و شاخص تحمل تنش (STI) در غلظت ۳۰۰ میلی گرم در لیتر کادمیوم، بهتر از گروه اول عمـل کرد. در نتیجـه گروه دوم بهعنوان گروه برتر انتخاب شد. در میان جدایههای گره در لیتر کادمیوم، بهتر از گروه اول عمـل کرد. در نتیجـه گروه بالاترین آSTI (به ترتیب ۱۹۷۸) در غلظت ۳۰۰ میلی گرم در لیتر کادمیوم، بهتر از گروه اول عمـل کرد. در نتیجـه گروه دوم بهعنوان گروه برتر انتخاب شد. در میان جدایههای گره در لیتر کادمیوم و همچـنین بالاترین تولید زیست وده (۱۰، ۲۰۷ دوم به عنوان گروه در لیتر) را در غلظتهای صفر و ۲۰۰ میلی گرم در لیتر کادمیوم و همچـنین بالاترین تولید زیست وده (۲۰ زیست وده (۶۶/۰ گرم در لیتر) را در مطلت ۳۰۰ میلی گرم در لیتر کادمیوم و همچـنین بالاترین تولید (۱۰ تولید زیست وده (۶۶/۰ گرم در لیتر) را در مطلت ۳۰۰ میلی گرم در لیتر کادمیوم و همچـنین بالاترین آ

كلمات كلیدی: تجزیه خوشهای، قارچ، سمیت عنصر سنگین، نرخ مهار، شاخص تحمل تنش

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